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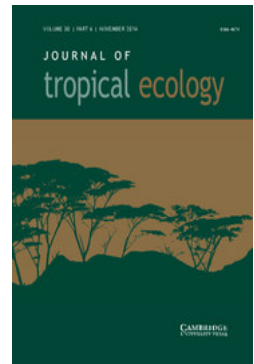
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Distribution of the community of frugivorous birds along a successional gradient in a tropical dry forest in south-western Mexico

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Abstract: Few reports have described the relationship between the distribution of frugivorous birds and vegetation successional changes in dry forests. We assessed the abundance and behaviour of frugivorous birds in early, intermediate and mature dry forests in the Balsas river basin, Guerrero, Mexico. We selected nine dry-forest fragments, three fragments per stage, in these three stages of succession. We analysed the vegetation, estimated bird abundances in 10-min count periods, and recorded the way birds process fruits in circular plots (11–15 plots per fragment, 123 plots in total). Birds were classified as seed predators (15% of all individuals in this study), pulp consumers (15%) or legitimate dispersers (70%). Bird abundance was higher in mature forests in the dry season, while abundance and richness of legitimate dispersers and seed predators were positively related to vegetation complexity. Mature forests have a high vegetation complexity and a high cover of *Bursera* species that produce fruit during the dry season. During the rains, abundance was higher in early-successional sites when the zoochorous plants produced fruit. Legitimate disperser migrants (i.e. *Tyrannus vociferans*, *Myiarchus cinerascens* and *M. tyrannulus*) were widespread, helping the establishment of zoochorous trees such as *Bursera* spp. in early-successional forests.

Key Words: *Bursera*, disturbance, forest succession, migratory birds, tyrannids, zoochorous trees

INTRODUCTION

Agriculture and cattle ranching have reduced the original vegetation to patches of different successional stages of tropical dry forests (Portillo-Quintero & Sánchez-Azofeita 2010, Quesada *et al.* 2009). The size of the remaining fragments may influence bird species richness (Maldonado-Coelho & Marini 2000), but the composition and structure of the vegetation remaining on the original patches may have a stronger effect on the number of bird species than the size of the fragment (Maldonado-Coelho & Marini 2000, Stouffer & Bierregaard 1995). Bird richness may also be high in intermediate stages due to the presence of species from early and mature forests (Connell 1978).

In addition to the arrival of migratory birds at the end of the rainy season, birds respond to the physiognomic

changes of the vegetation and to the available food resources. For instance, *Bursera* trees, a dominant component of mature dry forests of western Mexico (Rzedowski *et al.* 2005), produce fruit during the dry season. There is ample evidence that frugivores, such as flycatchers and vireos, include *Bursera* drupes as an important part of their diet (Ramos-Ordoñez & Arizmendi 2011). These birds also consume arthropods and fruit from other plants, but in the dry season some species seem to prefer this resource to the sugary fruit produced by other plants.

Flycatchers in particular remove large amounts of fruit from these trees (Ramos-Ordoñez & Arizmendi 2011). In addition, migratory birds move readily between mature and early-successional patches (Hutto 1998, Smith *et al.* 2001). These habitat shifts facilitate the dispersal of seeds and the establishment of woody vegetation in disturbed environments (Galindo-González *et al.* 2000, Tellería *et al.* 2005).

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Few studies have focused on the relationship between the distribution of frugivorous birds and habitat changes in dry forests (Chazdon *et al.* 2011, Ortiz-Pulido *et al.* 2000, Smith *et al.* 2001). We assessed the influence of the vegetation structure, successional stage and season on the diversity and distribution of frugivorous birds in tropical dry forest in south-western Mexico. Since early-succession sites were larger than mature and intermediate patches, we also assessed the influence of patch size on the number of frugivorous bird species. We hypothesized that: (1) Frugivorous bird richness should respond positively to the complexity of the vegetation. (2) Species richness and abundance of frugivores should be higher in late-successional stages during the dry season because the production of fruit by zoochorous trees such as *Bursera* spp., a preferred food source, is higher during this time of the year. Migratory species, on the other hand, are ubiquitous (Hutto 1998) and their distribution would be widespread. (3) Species richness should be higher in transitional succession stages during the rains, according to the intermediate disturbance hypothesis (Connell 1978).

METHODS

Study area and sampling sites

The study was conducted in the area known as La Organera–Xochipala, 18°03′46.65″–17°42′11.14″N, 99°36′36.50″–99°35′30.46″W. The region belongs to the biotic province of the Balsas river basin, in the Mexican state of Guerrero (Figure 1). The orography in the region is rugged, with an average altitude of 1100 m asl. The predominant climate is semi-dry, with an annual average temperature of 23.9 °C, and an annual rainfall of 684 mm (Peralta 1995). Rain falls mainly from June to October (Meza & López 1997).

We studied nine dry-forest fragments in three stages of succession (i.e. seral stages) that had been mostly unmanaged for different periods of time since the last major disturbance (i.e. clear-cutting and fire): (1) Three disturbed sites (early stage of succession (ES) ~20 y) consisting of native vegetation in land used for cattle ranching and, to a lesser extent, seasonal agriculture. Today the vegetation continues to be felled for wood. The sites have a low slope and there are only scattered trees on rocky ground that is not suitable for large-scale agriculture (Almazán-Núñez *et al.* 2012). (2) Three sites within the transition between a mature forest and fragmented areas (intermediate stage (IS) ~35 y). The matrix surrounding the sites consisted of pasture and corn and bean seasonal fields. These sites have a steeper slope which varies from 15° to 30°, and have been used for seasonal corn production as well as cattle ranching. There

are still some patches with eroded soils. Nonetheless, to a large extent, they have developed structural and floristic elements from the original vegetation. (3) Three sites with a closed canopy (mature stage (MS) > 50 y), characterized by the presence of a tree cover that is typical of mature dry forest (e.g. dominance of the *Bursera* spp.). The slope here is > 25° and therefore these areas have not been used on a large scale. In past years all sites were clear-cut and burned to open areas for agricultural and/or ranching activities. This information was obtained from interviews with farmers.

The area of the sites ranged from 18 to 181 ha. The mean distance between sites was 0.80 km with a range of 0.20 km (MS1 to ES3) to 1.78 km (ES1 to MS1) (Figure 1).

Habitat structure

We established 41 30-m diameter (0.28 ha) circular plots (two sites with 15 plots each, one with 11 plots) for each successional stage (three sites per stage; Figure 1). In each of these ($n = 123$ plots) we defined two perpendicular transects pointing to the four cardinal points. We recorded and identified every individual tree and shrub (dbh ≥ 10 cm) that intersected the lines (Almazán-Núñez *et al.* 2012). The crown size or cover of each plant was estimated with the formula for an ellipse using the maximum and minimum diameter lengths (Muller-Dombois & Ellenberg 1974). We also counted the number of individual trees and shrubs (shrubs were differentiated from trees by their relative short size (< 6 m) and multiple stems) in each plot (density). Stratification of the vegetation was determined with an optical square marked with two perpendicular axes (Montaña & Ezcurra 1980). In each plot we recorded height and number of times the foliage touched the point of intersection of the two axes. This procedure was repeated every 50 cm (60 times) oriented to the four cardinal directions in each transect. Heights were grouped in 1-m intervals and foliage height diversity (FHD) was estimated with the Shannon–Wiener index.

In addition to the stratification of the vegetation, 15 additional variables were assessed: (1) shrub density, (2) shrub cover, (3) density of *Bursera* spp. trees, (4) cover of *Bursera* spp. trees, (5) density of succulent species (i.e. Cactaceae plants), (6) cover of succulent species, (7) density of other zoochorous species (i.e. species whose fruits are eaten by birds, but do not belong to *Bursera* or Cactaceae), (8) cover of zoochorous species, (9) total plant density with a dbh ≥ 10 cm, (10) total cover of every plant with a dbh ≥ 10 cm, (11) diameter at breast height (dbh ≥ 10 cm), (12) mean height of shrub and trees species, (13) shrub species richness, (14) tree species richness, and (15) total species richness (trees and

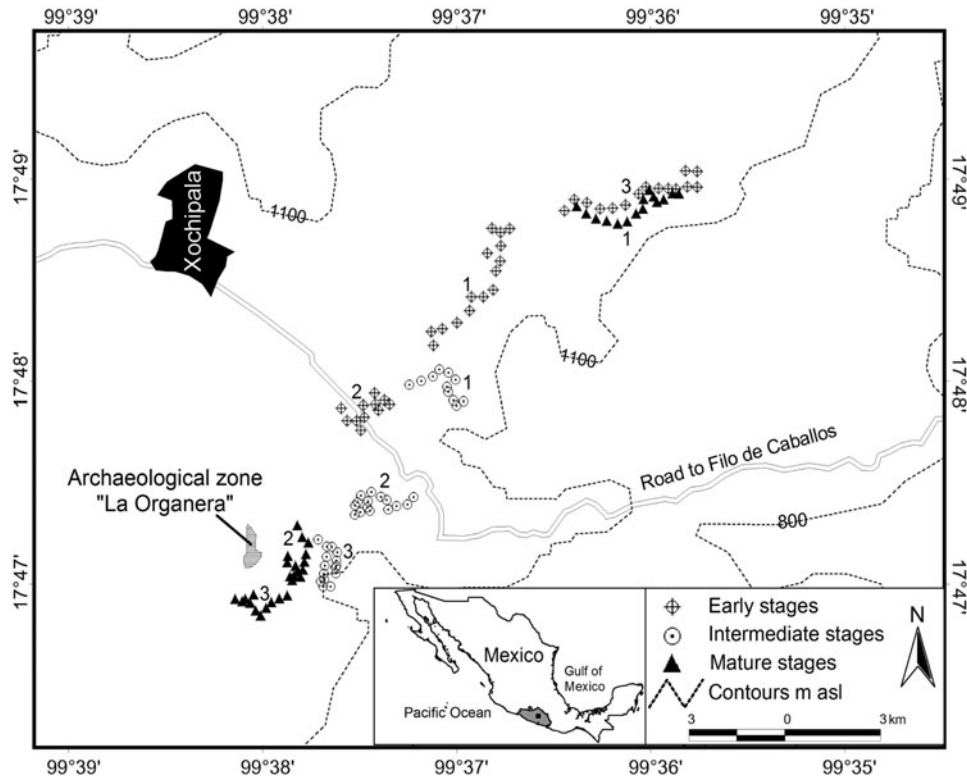


Figure 1. Map of the study region in the state of Guerrero, southern Mexico, showing the sampling plots in La Organera–Xochipala tropical dry forest. The numbers (1–3) represent the three sites of each successional stage.

shrubs). Detailed vegetation analyses for these sites are described in Almazán-Núñez *et al.* (2012).

Bird census

Bird censuses took place from February 2010 to February 2011 and we used a fixed-radius point count method with a 30-m radius. This distance has been recommended for tropical dry forests (Hutto *et al.* 1986). Each point, established in the same plots in which the vegetation had been characterized, was separated by approximately 200 m to avoid counting the same bird more than once. Every frugivorous bird seen or heard within the point count radius was recorded. Observations took place during the birds' period of highest activity: mornings (6h30–11h00) and afternoons (16h30–19h00). All individuals seen or heard within a 10-min period at each point were recorded. This time lapse is long enough to allow most birds present, including rare species, to be counted, and short enough so that the probability of counting the same bird more than once is minimized (Reynolds *et al.* 1980). Each site was visited eight times during the study period. Since the dry season is longer (7 mo) than the wet period (5 mo), plots were visited five times during the dry months and three during the rains. The abundance of each species was the average number resulting from all visits in each season.

We recorded the name of the species, as well as the number of individuals. We also observed, when possible, the way in which birds processed fruits and grouped them as potential legitimate dispersers (birds that eat the entire fruit and regurgitate or defecate the seeds further away from the plant from which the fruit was obtained), pulp consumers (species that consume the pulp and discard the seeds, usually under the parent plant), or seed predators (birds that break the seeds when eating them or eat the whole fruit and digest the seeds together with the pulp) (Jordano 1992, 1995; Traveset 1994). Because frugivorous birds can be both dispersers of some plant species and predators or pulp consumers of others (Herrera 2002), we considered the general interaction pattern (i.e. relationship between seed size and manner of processing; Levey 1987) based on the information gathered in the study areas, and particularly on the way in which they processed *Bursera* drupes. *Bursera* drupes were the main fruit frugivores ate and the trees of the genus are dominant in the study area (Almazán-Núñez *et al.* 2012). No direct foraging observations were recorded for some birds (for example, different species of dove) but, since they are known to be frugivores, we used the information provided in the Neotropical bird literature (Skutch 1983) to categorize them. Most of our observations coincided with the literature with the exception of species such as *Passerina versicolor*, *P. leclancherii* and *Spinus psaltria*,

which are generally categorized as granivores, but were seen eating the pseudo-aril of *Bursera* fruits. Since usually they discarded the seed of the drupes, we classified them as pulp eaters. Species were also separated according to seasonal status (i.e. local and latitudinal migrants and resident species).

Statistical analyses

Since the vegetation variables were highly intercorrelated, and it was difficult to assess which of them explained the bird species distribution, we used Principal Component Analyses (PCA) with varimax rotation for the 16 untransformed habitat-related variables to synthesize the structure of the vegetation. Furthermore, since birds probably recognize their habitats based on the simultaneous assessment of different vegetation characteristics (the niche gestalt suggested by James (1971)), it seems that this approach may be a more realistic approximation of the habitat space used by birds than the individual variables.

The Chao2 estimator was used to assess the comprehensiveness of the bird species survey. This non-parametric species richness estimator has been found to give reasonable approximations at small-grain-size sample units (i.e. circular plots; González-Oreja *et al.* 2010, Hortal *et al.* 2006). In addition, empirical results have also shown it is less dependent on sampling intensities than other approximation methods (Colwell & Coddington 1994, Hortal *et al.* 2006). Calculations were done with the software EstimateS version 8. Dominance was obtained with the Simpson index (Magurran 1988). Differences in abundance and dominance of birds between sites and successional stages during the dry and rainy seasons were compared with unbalanced two-way analysis of variance (ANOVA, factor 1: successional stage, factor 2: sites). After inspecting the statistical distribution, data were transformed to $\log(x + 1)$ in order to comply with the normality and homoscedasticity assumptions. Multiple comparisons were carried out using Tukey HSD test with SPSS v. 17.0. Since the sampling effort was different between sites, we used rarefaction to compare richness. The effect of vegetation on rarefied species richness, abundance and dominance of frugivorous birds was calculated with linear regressions. We also used regression analyses to assess the influence of the fragment size on the number of frugivorous bird species. All analyses were considered statistically significant when $P \leq 0.05$.

The relationship between species distribution and the vegetation principal components was analysed using a Canonical Correspondence Analysis (CCA). This technique is used to obtain synthetic environmental gradients of groups of ecological data (ter Braak & Verdonschot 1995) and to detect species distribution patterns that can be explained by a group of

environmental variables (Storch *et al.* 2002). Bird species with \leq three individuals for any given season were excluded from the analysis of that particular period. We tested the significance of the results with Monte Carlo permutations (999 simulations, $P \leq 0.05$). In this case, two statistical tests were used, one based on the first ordination axis, and the other based on the addition of all canonical axes (ter Braak & Verdonschot 1995). These analyses were run with the CANOCO 4.5 and CANODRAW 4.0 programs[©].

RESULTS

Habitat structure

The vegetation structure differed between successional stages. Tree density, vegetation cover and foliage height diversity were highest in the intermediate seral stage. This was because there were more young trees in different layers in these sites compared with mature forests, which had fewer but larger trees. Trees in the early-successional sites were more scattered than in the other fragments.

The structure of the vegetation was synthesized into three principal components that explained 89% of the variance (Table 1). The first component (PC1) explained 60% of the variance, and represented the complexity of the vegetation due to the high scores of variables such as vegetation height, foliage height diversity, density and cover of *Bursera*, cover and number of individual plants per plot tree richness, and total richness. The higher positive scores of the second component (19% of the explained variance) corresponded to the cover of other zoochorous plants which are present in less complex forest (i.e. *Ficus* spp. and *Capparis* spp.) and that did not include species of Burseraceae or Cactaceae and negatively to shrub richness, density and cover. The third component (9% of the explained variance) was related to the cover of succulent species of Cactaceae.

Composition of frugivorous birds

We recorded 31 species of frugivorous bird belonging to 13 families in the nine study sites (Appendix 1). Species like *Ortalis poliocephala*, *Leptotila verreauxi*, *Zenaidura macroura*, *Aratinga canicularis*, *Melanerpes chrysogenys*, *Myiarchus tyrannulus*, *M. cinerascens*, *Tyrannus verticalis*, *T. vociferans*, *Passerina leclancherii* and *Icterus pustulatus* were present in every site. Other species were found only in intermediate or mature stages of succession with higher tree cover, like *Momotus mexicanus* and *Myiarchus nuttingi*. There were no species found exclusively in early stages, although the four species that feed on pulp were more abundant in sites in early stages of succession. We also found six species (15% of all the

Table 1. Ordination (PCA) values of the first three factors for habitat variables in nine sites in a tropical dry forest in south-western Mexico. The sites represented different successional stages. Significant correlations are highlighted with * $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$.

	PC1	PC2	PC3
Floristic variables			
Shrub richness	0.27	-0.73*	-0.47
Tree richness	0.86**	0.43	0.09
Total richness	0.94***	-0.13	-0.25
Structural variables			
Shrub density (ind ha ⁻¹)	-0.43	-0.88**	-0.01
<i>Bursera</i> density (ind ha ⁻¹)	0.93***	0.22	0.23
Succulent species density (ind ha ⁻¹)	-0.56	-0.49	-0.24
Other zoochorous species density (ind ha ⁻¹)	0.25	0.93***	0.06
Total plant density (ind ha ⁻¹)	0.86**	0.31	0.20
Shrub cover (m ²)	-0.02	-0.86**	0.25
<i>Bursera</i> cover (m ²)	0.90***	0.23	0.34
Succulent species cover (m ²)	0.40	-0.15	0.89**
Other zoochorous species cover (m ²)	0.33	0.82**	0.08
Total plant cover (m ²)	0.90***	0.09	0.34
Tree diameter at breast height (cm)	0.56	0.66*	0.44
Mean height of all plants (m)	0.73*	0.62	0.03
Foliage height diversity (H')	0.73*	0.31	0.44
Eigenvalues	9.65	3.07	1.49
Explained variance (%)	60	19	9
Cumulative variance (%)	60	79	89

individuals) that predated seeds, and 21 legitimate potential dispersers (70% of the total; Appendix 1). On average we recorded 90% of the total number of expected frugivorous according to Chao2 during the dry season and 93% during the rainy season (Appendix 2).

Bird-habitat relationships

In the rainy season, the rarefied number of species (MaoTau) was lower in the early-successional sites in comparison with intermediate and mature stages. In the dry season the number of species was similar between successional stages (Appendix 2).

In the dry season, intermediate and mature-forest sites had more individuals than those of early stages ($F = 35$, $P < 0.001$; Figure 2a). Nevertheless, the interaction between sites and stages was significant ($F = 5.09$, $P < 0.01$; Table 2), because one of the early-successional sites (ES2) did not have a significant difference with regards to one of the mature-forest sites (MS1). In the rainy season, the number of individuals was different between sites but not between seral stages ($F = 5.31$, $P < 0.01$; Figure 2b). One of the early-successional sites (ES2) had the highest abundance of birds in this season; the site with the lowest abundance was MS1 (Table 2).

During the dry season, dominance of the frugivorous birds was consistently higher in the early stages ($F = 18.9$, $P < 0.001$; Figure 2c), mainly due to the presence of high numbers of pulp consumers such as *Haemorhous*

mexicanus and *Spinus psaltria*. In this case, the interaction between sites and successional stages was not significant (Table 2). In the rainy season, there were significant differences in the dominance of bird species between sites ($F = 14.6$, $P < 0.001$; Table 2), but not between successional stages ($F = 1.05$, $P > 0.05$; Table 2).

During the dry season, there was a negative relationship between the size of habitat fragments and the rarefied number of species ($F_{2,7} = 16.0$, $P = 0.005$), i.e. the smaller fragments had more bird species. The relationship between the size of the fragments and species richness was not significant for the rainy season ($F_{2,7} = 0.34$, $P = 0.57$). Abundance of frugivorous birds and that of migratory birds increased significantly with the vegetation complexity during the dry season (Figure 3).

The rarefied species richness and abundance of seed predators (SeedPred) were positively correlated with the first axis of the vegetation ordination (Figure 4a). Richness (Figure 4b) and abundance (Figure 4c) of potential legitimate dispersers (SeedDisp) also increased in sites with complex vegetation. Abundance of flycatchers, a group of birds that can remove large quantities of fruit (in particular from *Bursera* trees) in the region, showed the same tendencies (Figure 4c). The number of birds that feed on pulp (PulpCons) was negatively correlated with the first vegetation ordination axis (Figure 4d).

Although the Monte Carlo test for the first axis was not significant (F -ratio = 1.85, $P = 0.07$) in the dry season, it became significant with the addition of two other axes (F -ratio = 1.77, $P = 0.01$). The first axis of the CCA

Table 2. Two-way ANOVA comparisons of frugivorous bird diversity and abundance between three successional stages in a dry forest in south-western Mexico. * $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$.

	df	Dry season		Rainy season	
		Abundance	Dominance	Abundance	Dominance
Successional stage	2	35.0***	18.9***	0.37	1.05
Site	2	0.81	4.05*	5.31**	14.6***
Stage × Site	4	5.09**	0.34	0.57	1.00

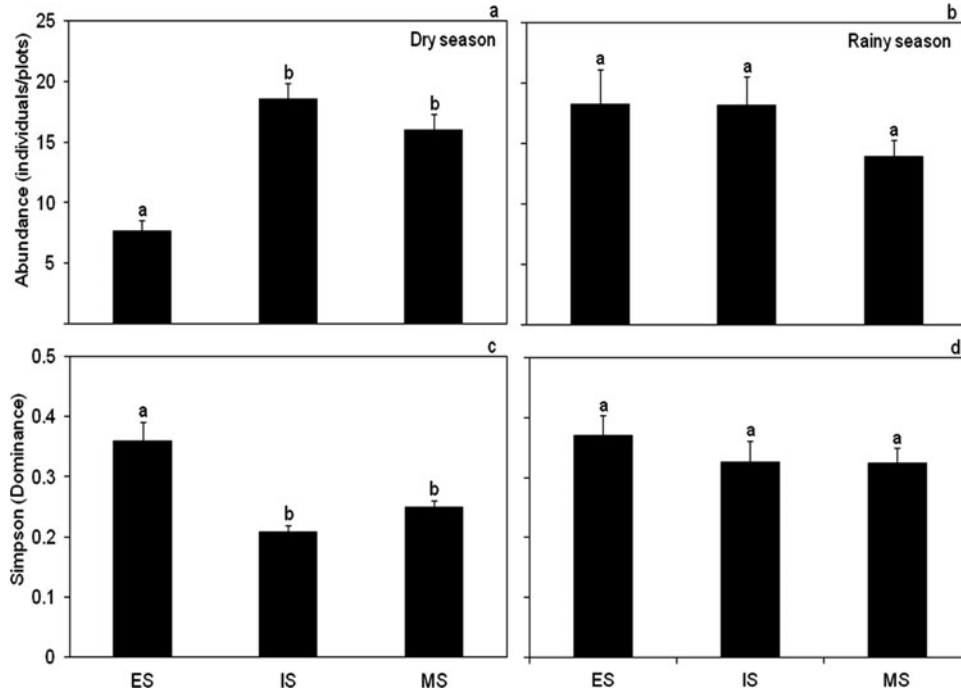


Figure 2. Mean abundance and dominance values for the frugivorous bird community in three different successional stages in a tropical dry forest in south-western Mexico during the dry (a, c) and rainy seasons (b, d). Different letters indicate significant differences ($P < 0.05$) according to Tukey HSD test. ES, early stage; IS, intermediate stage; MS, mature stage.

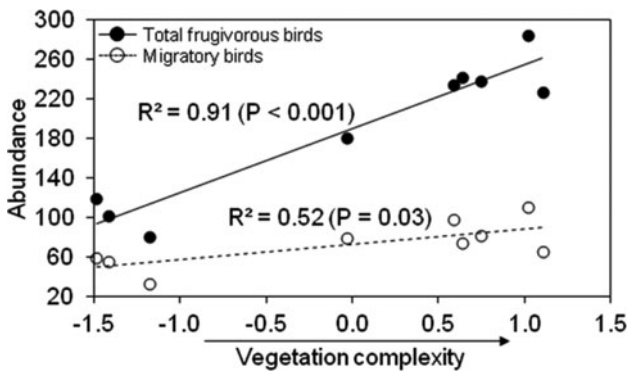


Figure 3. Relationships between abundance of migratory birds and total number of frugivorous birds, with the first factor of the PCA analysis of the vegetation in nine sites representing different successional stages in a south-western Mexico dry forest. The first ordination factor represented the early to advanced succession gradient.

(eigenvalue, $\lambda = 0.12$) explained 52.4% of the variation in the relationship between distribution of frugivorous birds and habitat variables, and had a positive correlation with the first vegetation axis (vegetation complexity = VegCom; -0.73 , $P < 0.05$). The other two components (succulent cover = SucCov and zoochorous species of disturbed sites = ZooDis), explained 31.3% and 16.7% of the variation, and were correlated with the second (-0.70 , $P < 0.05$) and third (-0.67 , $P < 0.05$) axes respectively.

The potential legitimate dispersers *Myiarchus nuttingi*, *Melanerpes chrysogenys*, *M. hypopoli*, *Myiodynastes luteiventris*, *Icterus wagleri* and the seed predators *Aratinga canicularis*, *Leptotila verreauxi* and *Pheucticus melanocephalus* were positively related to the first axis of the plant ordination (Figure 5a). Pulp consumers such as *Passerina versicolor*, *P. leclancherii* and *Spinus psaltria* were associated with the second ordination component,

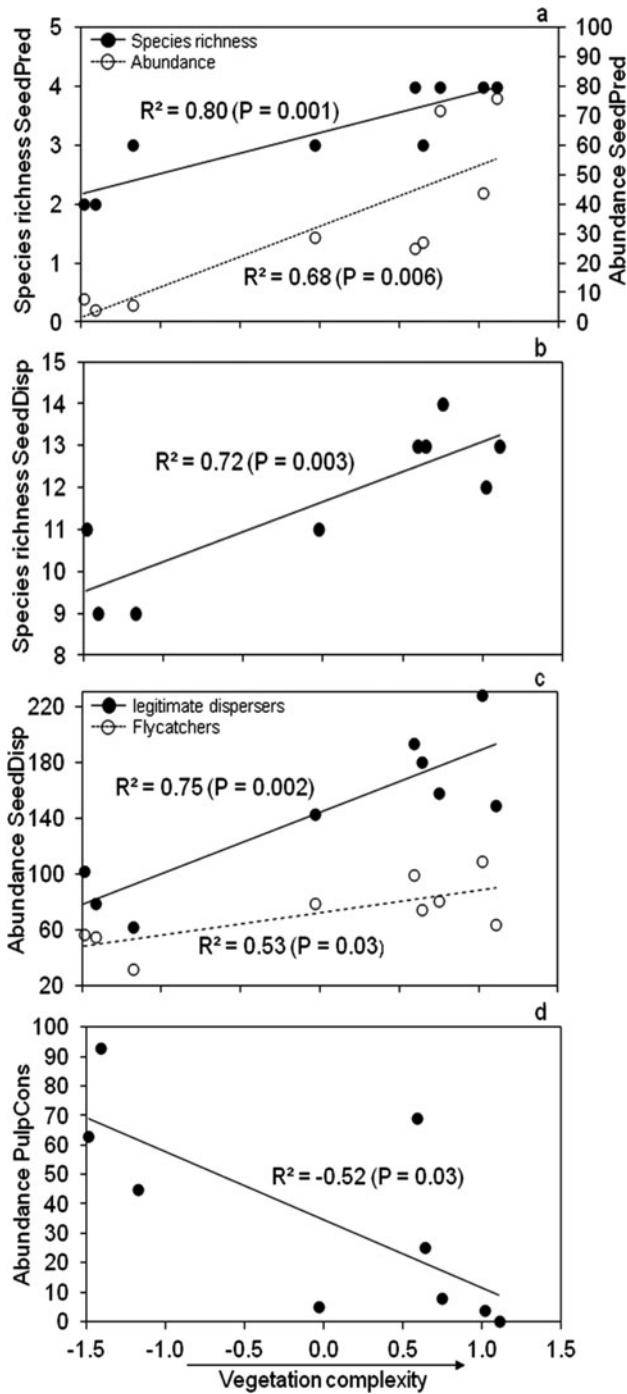


Figure 4. Relationships between different frugivorous bird guilds and the first factor of a PCA ordination of the vegetation in a dry forest in south-western Mexico. The first ordination factor represented the early to advanced successional gradient. During the dry season (a–c). During the rainy season (d). Abbreviations are indicated as SeedPred (seed predators), PulpCons (pulp consumers), SeedDisp (potential legitimate dispersers).

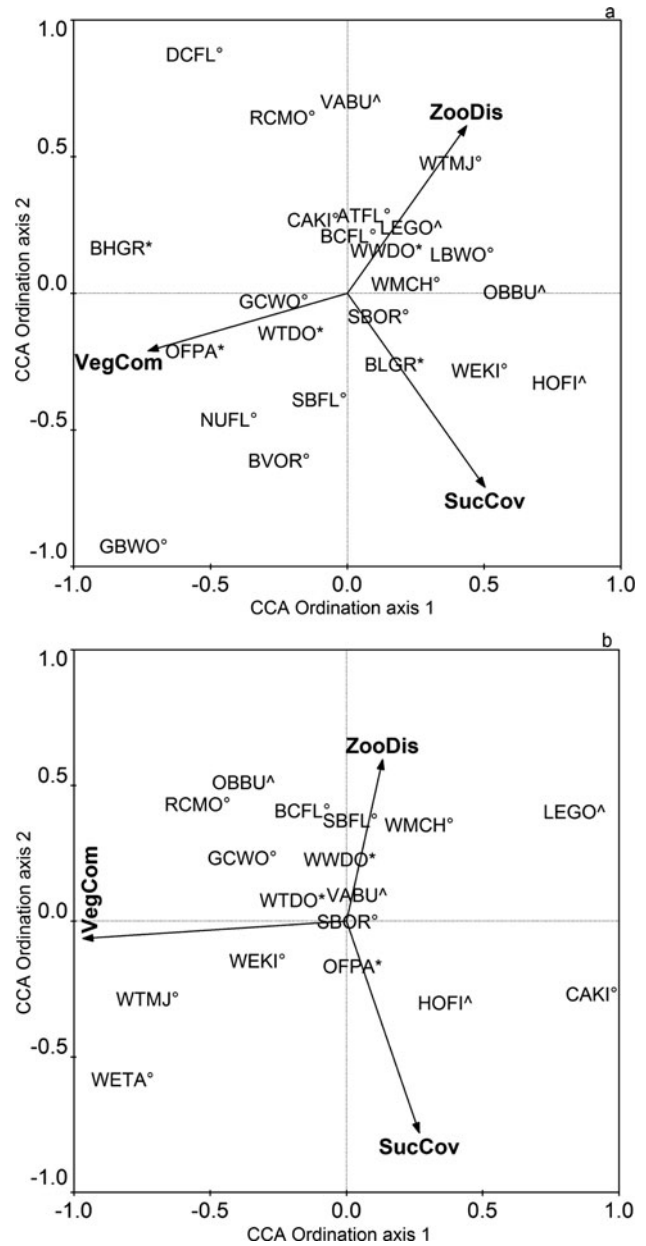


Figure 5. Canonical ordination of a frugivorous bird community in a dry forest in south-western Mexico during two contrasting seasons: dry season (a) and rainy season (b). Independent factors are the three main factors obtained from a PCA ordination of the vegetation. (°) potential legitimate dispersers, (*) seed predators, (ˆ) pulp consumers. Variables are indicated as: VegCom, vegetation complexity; ZooDis, zoochorous species of disturbed sites; SucCov, succulent cover. Abbreviations of bird species are given in Appendix 1. Species with \leq two individuals were excluded from the analyses.

as were the migratory legitimate dispersers *Tyrannus vociferans*, *Myiarchus cinerascens* and *M. tyrannulus*. The pulp consumer *Haemorhous mexicanus*, the potential legitimate disperser *T. verticalis*, and the seed predator *Passerina caerulea*, were associated with the cover of succulent plants (Figure 5a).

Monte Carlo permutation test was not significant for the first axis (F-ratio = 1.69, $P = 0.41$) nor for the total of canonical axis (F-ratio = 1.16, $P = 0.34$) in the rainy season. Nevertheless, the distribution of frugivorous groups showed a similar pattern as in the dry months. Legitimate migratory dispersers such as *Tyrannus verticalis* and *Piranga ludoviciana* and residents like *Calocitta formosa* and *Melanerpes chrysogenys*, as well as the seed predator *Leptotila verreauxi* (Figure 5b) were associated with the first vegetation axis (vegetation complexity). In contrast, the pulp consumers *Spinus psaltria* and *Passerina versicolor* and the legitimate disperser *Ortalis poliocephala*, were associated with the second axis (zoochorous species of disturbed sites). *Haemorhous mexicanus* was again associated with succulent cover.

DISCUSSION

Due to rapid changes in land use, sites represented by early successional stages generally occupy larger areas than forests in advanced succession (Quesada *et al.* 2009). Usually younger forests are structurally simple (Lasky & Keitt 2010). This may explain the negative correlation between species richness and size of the fragments during the dry season in this study. On the other hand, there was a positive correlation between species richness and vegetation complexity. This result supports our first hypothesis and agrees with other studies that have found more bird species in small fragments with high vegetation complexity (Echeverría *et al.* 2007, Gillespie & Walter 2001, Lasky & Keitt 2010, Maldonado-Coelho & Marini 2000, Newmark 1991, Stouffer & Bierregaard 1995). In addition, smaller fragments in our study sites had a larger number of species and cover of *Bursera* spp. (Almazán-Núñez *et al.* 2012). Drupes of *Bursera* are removed by different groups of frugivorous birds in the dry season (Greenberg *et al.* 1995, Ortiz-Pulido & Rico-Gray 2006, Ramos-Ordoñez & Arizmendi 2011).

Our results do not allow us to clarify which of the variables associated with axis 1 of the vegetation ordination could have a definite effect on the distribution of the birds. Nevertheless, the preference for the fruits of *Bursera* spp. could explain the high abundances and richness of legitimate dispersers, seed predators and migratory frugivores in the patches with a high vegetation complexity, as predicted by our second hypothesis. This fruit can be especially important to migratory species, because they provide energy which they store (Ramos-Ordoñez *et al.* 2012), improving their chances of survival during the migratory season, allowing them to return to their mating sites (Tellería *et al.* 2005).

In addition to food resources, resident legitimate dispersers (i.e. *Melanerpes chrysogenys* and *Myiarchus*

nuttingi) and seed predators (*Aratinga canicularis*, for example) may favour habitats with a complex vegetation, because they provide better opportunities for nesting and foraging (cactus species and taller trees) in the dry months (Ding *et al.* 2008, Palomera-García 2010). On the other hand, species such as *Spinus psaltria* and *Passerina leclancherii* have been associated with open areas with scattered trees (Howell & Webb 1995), and were more abundant in open fragments. These results coincide with reports by Chazdon *et al.* (2011), who found a higher dominance of species of these birds in disturbed areas of the Chamela-Cuixmala region in western Mexico.

During the wet season, the relationship between richness and size of the study sites was not significant. In addition, and contrary to our third hypothesis, we did not find a higher richness in the intermediate succession sites. The number of individuals, on the other hand, was highest in one of the early-successional sites. *Ficus cotinifolia* and *Capparis* spp., which are typical of disturbed vegetation, were abundant in that site. The two species produce abundant fruit that is consumed by frugivores during the rainy season (Durán-Ramírez *et al.* 2010, Lorea-Hernández 2004) and may have attracted large numbers of birds and pulp consumers in particular. Chazdon *et al.* (2011) reported similar results in the study they conducted in a dry forest in western Mexico.

The absence of *Bursera* drupes could explain the absence of *Myiarchus tuberculifer* and *Myiarchus tyrannulus* for the most part of the rainy season. Although some frugivorous birds can change the resources they use in response to seasonal changes (Carnicer *et al.* 2009), species such as the flycatchers mentioned above probably migrate locally when fruits are scarce, as has been reported for some other species of dry-forest flycatchers (Ramos-Ordoñez & Arizmendi 2011).

Even though the number of migrant birds was high in mature and intermediate stages, where the vegetation was more complex, it has been suggested that these species use indistinct patches of vegetation as intermediate stops in their migration routes (Hutto 1998). This was supported by the ordination results. The canonical correspondence analyses showed that *Tyrannus vociferans*, *Myiarchus cinerascens* and *M. tyrannulus* were ubiquitous in the dry season, in agreement with our second hypothesis. Migratory birds can therefore help to establish zoochorous trees in perturbed or early succession stages, increasing the connectivity between isolated fragments by facilitating the movement of seeds between patches in different stages (Bascompte & Jordano 2007, Tellería *et al.* 2005). As a consequence, migratory birds help in the recovery of woody vegetation (Galindo-González *et al.* 2000, Guevara & Laborde 1993, Herrera & García 2009).

In our study, plants of *Opuntia* spp. were common in early successional stages, while *Neoboubaunia*

mezcalaensis had high importance values in intermediate stages. These plant species had high scores in the third component (succulent cover) and have been associated with *Haemorhous mexicanus*, a sparrow that commonly feeds on fruit and nectar from cacti flowers (Godínez-Álvarez *et al.* 2002).

Conclusion

A decrease in mature vegetation in tropical dry forests means a decline in abundance and composition of frugivorous bird species and, therefore, a disruption in the interactions between frugivores and seed dispersal. Nevertheless, our results indicated that the vegetation of early stages of succession could provide an important habitat for some migratory species, which are known for eating large amounts of fruits (Bascompte & Jordano 2007, Ramos-Ordoñez & Arizmendi 2011). From this perspective, all stages of succession are important to maintain the interaction between plants and frugivores.

On the other hand, even though most of the frugivores were common in the three stages of succession, there were some species that were found only in mature stages (*Myiarchus tuberculifer*, *Pheucticus melanocephalus*) or which were more abundant in these sites (*Aratinga canicularis*, *Melanerpes chrysogenys*). Some traits of mature forests, such as climatic buffering and the presence of large trees can influence habitat choices and nesting sites of these resident species (Ding *et al.* 2008, Smith *et al.* 2001). If land-use tendencies in the region and, in the Balsas river basin in general, continue as they are today, these bird species can be seriously affected, and also the conservation of several *Bursera* species will be compromised, in particular those that are endemic to the Balsas basin (e.g. *B. longipes*, *B. chemapodicta*).

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LITERATURE CITED

- ALMAZÁN-NÚÑEZ, R. C., ARIZMENDI, M. C., EGUIARTE, L. E. & CORCUERA, P. 2012. Changes in composition, diversity and structure of woody plants in successional stages of tropical dry forest in southwest Mexico. *Revista Mexicana de Biodiversidad* 83:1096–1109.
- AOU (AMERICAN ORNITHOLOGISTS' UNION). 1998. *Check-list of North American birds*. (Seventh edition). American Ornithologists' Union, Washington, DC. 829 pp.
- BASCOMPTE, J. & JORDANO, P. 2007. Plant-animal mutualistic networks: the architecture of biodiversity. *Annual Review of Ecology, Evolution and Systematics* 38:567–593.
- CARNICER, J., JORDANO, P. & MELIÁN, C. J. 2009. The temporal dynamics of resource use by fruit-eating birds: a network approach. *Ecology* 90:1958–1970.
- CHAZDON, R. L., HARVEY, C. A., MARTÍNEZ-RAMOS, M., BALVANERA, P., STONER, K. E., SCHONDUBE, J. E., AVILA-CABADILLA, L. D. & FLORES-HIDALGO, M. 2011. Seasonally dry tropical forest biodiversity and conservation value in agricultural landscapes of Mesoamerica. Pp. 195–219 in Dirzo, R., Young, H. S., Mooney, H. A. & Ceballos, G. (eds.). *Seasonally dry tropical forest: ecology and conservation*. Island Press, Washington, DC.
- COLWELL, R. K. & CODDINGTON, J. A. 1994. Estimating terrestrial biodiversity. *Philosophical Transactions of the Royal Society B: Biological Sciences* 345:101–118.
- CONNELL, J. H. 1978. Diversity in tropical rain forests and coral reefs. *Science* 199:1302–1310.
- DING, T. S., LIAO, H. C. & YUAN, H. W. 2008. Breeding bird community composition in different successional vegetation in the montane coniferous forest zone of Taiwan. *Forest Ecology and Management* 255:2038–2048.
- DURÁN-RAMÍREZ, C. A., FONSECA-JUÁREZ, R. M. & IBARRA-MANRÍQUEZ, G. 2010. Estudio florístico de *Ficus* (Moraceae) en el estado de Guerrero, México. *Revista Mexicana de Biodiversidad* 81:239–262.
- ECHEVERRÍA, C., NEWTON, A. C., LARA, A., BENAYAS, J. M. R. & COOMES, D. A. 2007. Impacts of forest fragmentation on species composition and forest structure in the temperate landscape of southern Chile. *Global Ecology and Biogeography* 16:426–439.
- GALINDO-GONZÁLEZ, J., GUEVARA, S. & SOSA, V. J. 2000. Bat- and bird-generated seed rains at isolated trees in pastures in a tropical rainforest. *Conservation Biology* 14:1693–1703.
- GILLESPIE, T. W. & WALTER, H. 2001. Distribution of bird species richness at a regional scale in tropical dry forest in Central America. *Journal of Biogeography* 28:651–662.
- GODÍNEZ-ÁLVAREZ, H. O., VALIENTE-BANUET, A. & ROJAS-MARTÍNEZ, A. 2002. The role of seed dispersers in the population dynamics of the columnar cactus. *Neobuxbaumia tetetzo*. *Ecology* 83:2617–2629.
- GONZÁLEZ-OREJA, J. A., DE LA FUENTE, A. A., HERNÁNDEZ-SANTÍN, L., BUZO-FRANCO, D. & BONACHE-REGIDOR, C. 2010. Evaluación de estimadores no paramétricos de la riqueza de especies. Un ejemplo con aves en áreas verdes de la ciudad de Puebla, México. *Animal Biodiversity and Conservation* 33:31–45.
- GREENBERG, R., FOSTER, M. S. & MÁRQUEZ-VALDELAMAR, L. 1995. The role of the white-eyed vireo in the dispersal of *Bursera* fruit on the Yucatan Peninsula. *Journal of Tropical Ecology* 11:619–639.

- GUEVARA, S. & LABORDE, J. 1993. Monitoring seed dispersal at isolated standing trees in tropical pastures: consequences for local species availability. *Vegetatio* 107/108:319–338.
- HERRERA, C. M. 2002. Seed dispersal by vertebrates. Pp. 185–208 in Herrera, C. M. & Pellmyr, O. (eds.). *Plant animal interactions: an evolutionary approach*. Blackwell Scientific Publications, Oxford.
- HERRERA, J. M. & GARCÍA, G. 2009. The role of remnant trees in seed dispersal through the matrix: being alone is not always so sad. *Biological Conservation* 142:149–158.
- HORTAL, J., BORGES, P. A. & GASPAR, C. 2006. Evaluating the performance of species richness estimators: sensitivity to sample grain size. *Journal of Animal Ecology* 75:274–287.
- HOWELL, S. N. G. & WEBB, S. 1995. *A guide to the birds of Mexico and northern Central America*. Oxford University Press, New York, 849 pp.
- HUTTO, R. L. 1998. On the importance of stopover sites to migrating birds. *Auk* 115:823–825.
- HUTTO, R. L., PLETSCHE, S. M. & HENDRICKS, P. 1986. A fixed-radius point count method for nonbreeding and breeding season use. *Auk* 103:593–602.
- JAMES, F. C. 1971. Ordinations of habitat relationships among breeding birds. *Wilson Bulletin* 83:215–236.
- JORDANO, P. 1992. Fruits and frugivory. Pp. 105–156 in Fenner, M. (ed.). *Seeds: the ecology of regeneration in natural plant communities*. CABI, Wallingford.
- JORDANO, P. 1995. Spatial and temporal variation in the avian-frugivore assemblage of *Prunus mahaleb*: patterns and consequences. *Oikos* 71:479–491.
- LASKY, J. R. & KEITT, T. H. 2010. Abundance of Panamanian dry-forest birds along gradients of forest cover at multiple scales. *Journal of Tropical Ecology* 26:67–78.
- LEVEY, D. J. 1987. Seed size and fruit-handling techniques of avian frugivores. *American Naturalist* 129:471–485.
- LOREA-HERNÁNDEZ, F. G. 2004. Capparaeae. *Flora del Bajío y de Regiones Adyacentes*. 130:1–31.
- MAGURRAN, A. E. 1988. *Ecological diversity and its measurement*. Cambridge University Press, Cambridge. 192 pp.
- MALDONADO-COELHO, M. & MARINI, M. A. 2000. Effects of forest fragment size and successional stage on mixed-species bird flocks in southeastern Brazil. *Condor* 102:585–594.
- MEZA, L. & LÓPEZ, J. L. 1997. Vegetación y mesoclima de Guerrero. Pp. 1–53 in Diego-Pérez, N. & Fonseca, R. M. (eds.). *Estudios florísticos en Guerrero*. Facultad de Ciencias, UNAM, Mexico, D.F.
- MONTAÑA, C. & EZCURRA, E. 1980. Simple instrument for quick measurement of crown projections. *Journal of Forestry* 78:699.
- MULLER-DOMBOIS, D. & ELLENBERG, H. 1974. *Aims and methods of vegetation ecology*. John Wiley and Sons, New York. 547 pp.
- NEWMARK, W. D. 1991. Tropical forest fragmentation and the local extinction of understory birds in the eastern Usambara mountains, Tanzania. *Conservation Biology* 5:67–78.
- ORTIZ-PULIDO, R. & RICO-GRAY, V. 2006. Seed dispersal of *Bursera fagaroides* (Burseraceae): the effect of linking environmental factors. *Southwestern Naturalist* 51:11–21.
- ORTIZ-PULIDO, R., LABORDE, J. & GUEVARA, S. 2000. Frugivoría por aves en un paisaje fragmentado: consecuencias en la dispersión de semillas. *Biotropica* 32:473–488.
- PALOMERA-GARCÍA, C. 2010. Habitat use and local harvesting practices of the Orange-fronted Parakeet (*Aratinga canicularis*) in western Mexico. *Studies on Neotropical Fauna and Environment* 45:139–147.
- PERALTA, S. 1995. Cañón del Zopilote (área Papalotepec). Pp. 1–37 in Diego-Pérez, N. & Fonseca, R. M. (eds.). *Estudios florísticos en Guerrero*. Facultad de Ciencias, UNAM, Mexico, D.F.
- PORTILLO-QUINTERO, C. & SÁNCHEZ-AZOFEITA, G. A. 2010. Extent and conservation of tropical dry forest in the Americas. *Biological Conservation* 143:144–155.
- QUESADA, M., SÁNCHEZ-AZOFEITA, G. A., ÁLVAREZ-AÑORVE, M., STONER, K. E., AVILA-CABADILLA, L., CALVO-ALVARADO, J., CASTILLO, A., ESPÍRITO-SANTO, M. M., FAGUNDES, M., FERNANDES, G. W., GAMON, J., LOPEZARAIZA-MIKEL, M., LAWRENCE, B., CERDEIRA, L. P., POWERS, J. S., NEVES, F., ROSAS-GUERRERO, V., SAYAGO, R. & SÁNCHEZ-MONTOYA, G. 2009. Succession and management of tropical dry forests in the Americas: review and new perspectives. *Forest Ecology and Management* 258:1014–1024.
- RAMOS-ORDOÑEZ, M. F. & ARIZMENDI, M. C. 2011. Parthenocarpy, attractiveness and seed predation by birds in *Bursera morelensis*. *Journal of Arid Environments* 75:757–762.
- RAMOS-ORDOÑEZ, M. F., ARIZMENDI, M. C. & MÁRQUEZ-GUZMÁN, J. 2012. The fruit of *Bursera*: structure, maturation and parthenocarpy. *AoB PLANTS* doi: 10.1093/aobpla/pls027.
- REYNOLDS, S. T., SCOTT, J. M. & NUSSBAUM, R. A. 1980. A variable circular-plot method for estimating bird numbers. *Condor* 82:309–313.
- RZEDOWSKI, J., MEDINA, R. & CALDERÓN, G. 2005. Inventario del conocimiento taxonómico, así como de la diversidad y del endemismo regionales de las especies mexicanas de *Bursera* (Burseraceae). *Acta Botánica Mexicana* 70:85–111.
- SKUTCH, A. F. 1983. *Birds of tropical America*. University of Texas Press, Austin. 305 pp.
- SMITH, A. L., SALGADO-ORTIZ, J. & ROBERTSON, R. J. 2001. Distribution patterns of migrant and resident birds in successional forests of the Yucatan peninsula, Mexico. *Biotropica* 33:153–170.
- STORCH, D., GASTON, K. J. & CEPÁK, J. 2002. Pink landscapes: 1/f spectra of spatial environmental variability and bird community composition. *Proceedings of the Royal Society London B* 269:1791–1796.
- STOUFFER, P. C. & BIERREGAARD, R. O. 1995. Use of Amazonian forest fragments by understory insectivorous birds. *Ecology* 76:2429–2445.
- TELLERÍA, J. L., RAMÍREZ, A. & PÉREZ-TRIS, J. 2005. Conservation of seed-dispersing migrant birds in Mediterranean habitats: shedding light on patterns to preserve processes. *Biological Conservation* 124:493–502.
- TER BRAAK, C. J. F. & VERDONSHOT, P. F. M. 1995. Canonical correspondence analysis and related multivariate methods in aquatic ecology. *Aquatic Sciences* 57:255–289.
- TRAVESET, A. 1994. Influence of type of avian frugivory on the fitness of *Pistacia terebinthus*. *Evolutionary Ecology* 8:618–627.

Appendix 1. Species composition, seasonal status and relative abundance of frugivorous bird species in nine sites representing different successional stages in a tropical dry forest in south-western Mexico. Species names and systematic order follow AOU (1998), and supplements. The code after species name was used in the ordination analyses. Status: R (resident), M (migratory), LM (local migratory) according to Howell & Webb (1995) and personal observations. Frugivore type: PRED, Seed predator; DISP, potential legitimate disperser; CONS, pulp consumer.

Species	Status/frugivore type	Successional stage								
		Early			Intermediate			Mature		
		1	2	3	1	2	3	1	2	3
West Mexican chachalaca WMCH (<i>Ortalis poliocephala</i>)	R/DISP	0.27	0.45	1.00	1.55	0.73	0.87	0.20	0.53	0.36
White-tipped dove WTDO (<i>Leptotila verreauxi</i>)	R/PRED	0.60	0.55	1.00	1.09	1.07	2.20	1.07	1.33	2.73
White-winged dove WWDO (<i>Zenaida asiatica</i>)	M/PRED	0.07	0.36	0.27	0.18	0.73	0.60	0.07	0.33	0.09
Mourning dove MODO (<i>Z. macroura</i>)	M/PRED							0.07		
Orange-fronted parakeet OFPA (<i>Aratinga canicularis</i>)	R/PRED	1.07	0.55	0.40	1.45	1.40	2.20	1.20	3.67	5.18
Russet-crowned motmot RCMO (<i>Momotus mexicanus</i>)	R/DISP					0.80	0.60	0.20	0.40	0.64
Golden-cheeked woodpecker GCWO (<i>Melanerpes chrysogenys</i>)	R/DISP	0.67	0.64	1.40	4.64	4.60	4.87	2.80	4.80	5.64
Grey-breasted woodpecker GBWO (<i>M. hypopolius</i>)	R/DISP				0.09				0.07	0.27
Ladder-backed woodpecker LBWO (<i>Picoides scalaris</i>)	R/DISP	0.13	0.09			0.27	0.13		0.07	
Dusky-capped flycatcher DCFL (<i>Myiarchus tuberculifer</i>)	ML/DISP				0.18			0.33	0.47	
Ash-throated flycatcher ATFL (<i>M. cinerascens</i>)	M/DISP	0.67	0.36	0.80	1.00	0.60	0.80	1.13	0.67	1.00
Nutting's flycatcher NUFL (<i>M. nuttingi</i>)	M/DISP					0.07	0.07			0.09
Brown-crested flycatcher BCFL (<i>M. tyrannulus</i>)	ML/DISP	1.00	0.73	1.07	1.73	1.00	1.27	2.00	1.13	1.18
Sulphur-bellied flycatcher SBFL (<i>Myiodynastes luteiventris</i>)	M/DISP	0.33		1.47	0.55	0.33	0.27	0.33	0.33	1.00
Tropical kingbird TRKI (<i>Tyrannus melancholicus</i>)	R/DISP				0.09	0.07				
Cassin's kingbird CAKI (<i>T. vociferans</i>)	M/DISP	0.53	1.36	0.27	2.27	2.87	1.67	1.87	2.07	1.00
Western kingbird WEKI (<i>T. verticalis</i>)	M/DISP	4.40	4.73	0.67	4.36	4.47	7.20	1.20	4.07	7.36
Scissor-tailed flycatcher STFL (<i>T. forficatus</i>)	M/DISP									0.09
White-throated magpie-jay WTMJ (<i>Calocitta formosa</i>)	R/DISP	0.60		1.00	0.73	0.33	1.80	0.93	0.60	0.45
Golden vireo GOVI (<i>Vireo hypochryseus</i>)	R/DISP					0.07				
Warbling vireo WAVI (<i>V. gilvus</i>)	R/DISP						0.07			
Northern mockingbird NOMO (<i>Mimus polyglottos</i>)	R/DISP				0.09				0.07	
Blue grosbeak BLGR (<i>Passerina caerulea</i>)	R/PRED			0.07	0.09		0.07			0.09

Appendix 1. Continued.

Species	Status/frugivore type	Successional stage								
		Early			Intermediate			Mature		
		1	2	3	1	2	3	1	2	3
Varied bunting VABU (<i>P. versicolor</i>)	R/CONS		0.18	0.13	0.18	0.20	0.33	0.33	0.33	
Orange-breasted bunting OBBU (<i>P. leclancherii</i>)	R/CONS	0.33	0.64	0.27	0.27	0.67	0.07	0.07	0.47	0.18
Black-headed grosbeak BHGR (<i>Pheucticus melanocephalus</i>)	M/PRED								0.33	0.09
Western tanager WETA (<i>Piranga ludoviciana</i>)	M/DISP			0.07			0.07	0.07		0.27
Black-vented oriole BVOR (<i>Icterus wagleri</i>)	R/DISP	0.13	0.09			0.13			0.27	0.36
Streak-backed oriole SBOR (<i>I. pustulatus</i>)	R/DISP	1.93	4.55	1.93	3.27	3.73	4.00	2.07	3.00	4.18
Lesser goldfinch LEGO (<i>Spinus psaltria</i>)	R/CONS	2.07	1.91	2.93	0.09	1.53	0.40	0.47	0.60	
House finch HOFI (<i>Haemorhous mexicanus</i>)	R/CONS	2.40	7.45	0.07	7.91	1.60	0.27	0.27	0.20	

Appendix 2. Observed (Sobs), expected (Chao2) and rarefied (MaoTau) richness of the frugivorous bird community in nine sites representing three successional stages in a dry forest in south-western Mexico. ES, early stage; IS, intermediate stage; MS, mature stage.

Successional Stage	Sites	Dry season				Rainy season			
		Sobs	MaoTau	Chao2	Chao2%	Sobs	MaoTau	Chao2	Chao2%
ES	1	16	14	23	61.5	11	9	12	91.7
	2	14	13	15	77.8	13	9	15	86.7
	3	14	14	14	100.0	11	9	14	78.6
	1-3	20	20	20	100.0	15	15	15	100.0
IS	1	20	15	23	71.4	8	7	9	88.9
	2	20	15	21	83.3	15	12	21	71.4
	3	20	15	23	74.1	14	12	23	60.9
	1-3	26	21	29	76.5	17	16	20	85.0
MS	1	17	14	18	81.0	12	12	18	66.7
	2	21	15	25	70.0	16	14	25	64.0
	3	19	14	22	70.4	12	11	22	54.5
	1-3	23	20	24	92.0	20	20	21	95.2